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Chapter 1 : Human Health Water Quality Criteria and Methods for Toxics | Water Quality Criteria | US EPA

In particular, mercury analyses of water before the mid- to lates should be treated with suspicion. Download PDF sample Developing Ambient Water Quality Criteria for Mercury: A Probabalistic Site-Specific Approach.

How can I find the use designations for my specific water body of interest? Go to rules to of the OAC. Each of those rules covers a major drainage basin. Use the Water Body Use Designation Index on that page to find the rule number and page number of your water body of interest. Ohio drainage basins and associated rule numbers in OAC Beneficial use designations describe existing or potential uses of water bodies. They take into consideration the use and value of water for public water supplies, protection and propagation of aquatic life, recreation in and on the water, agricultural, industrial and other purposes. Ohio EPA assigns beneficial use designations to water bodies in the state. There may be more than one use designation assigned to a water body. Examples of beneficial use designations include: Attainment of use designations is based on specific numeric and narrative criteria. To ensure protection of those beneficial uses, Ohio EPA determines and assigns maximum concentrations for over chemicals. Narrative "free froms", located in rule of the OAC, are general water quality criteria that apply to all surface waters. These criteria state that all waters shall be free from sludge, floating debris, oil and scum, color and odor producing materials, substances that are harmful to human, animal or aquatic life, and nutrients in concentrations that may cause algal blooms. Very important components of this evaluation are the biological survey program and the biological criteria used to judge aquatic life use attainment. Numeric criteria are estimations of concentrations of chemicals and degree of aquatic life toxicity allowable in a water body without adversely impacting its beneficial uses. Although numeric criteria are applied to water bodies, they primarily are used to regulate dischargers through NPDES permits. Numeric criteria consist of chemical criteria, whole effluent toxicity levels and biological criteria. In these studies, organisms are exposed to known concentrations of a chemical under varying conditions. For aquatic life water quality criteria, the organisms exposed are a variety of fish, benthic macroinvertebrates and zooplankton. For human health water quality criteria, the organisms exposed are mammals, usually mice or rats. Based on these tests, guidelines or national criteria recommendations are established by U. Another class of chemical criteria are those associated with the Agricultural Water Supply use designation. These criteria protect against long term adverse effects on crops and livestock as a result of crop irrigation and livestock watering. Chemical water quality criteria are in Chapter of the OAC. A bioassay or toxicity test measures the degree of response of an exposed test organism to a specific chemical or effluent. WET can only be measured using living organisms, not by an instrument. WET consists of acute and chronic toxicity tests. Acute toxicity tests measure the responses of organisms that occur soon after exposure to a test substance. Chronic tests measure the long-term response to test substances. WET measures the accumulative effects of chemicals present in an effluent that cannot be assessed using chemical-specific criteria. Provisions addressing whole effluent toxicity are in rule 44 , rule , and paragraph B of rule of the OAC. Biological Criteria Biological criteria are based on aquatic community characteristics that are measured both structurally and functionally. These criteria are used to evaluate the attainment of aquatic life uses. The data collected in these assessments are used to characterize aquatic life impairment and to help diagnose the cause of this impairment. These three indices are based on species richness, trophic composition, diversity, presence of pollution-tolerant individuals or species, abundance of biomass, and the presence of diseased or abnormal organisms. Ohio EPA uses the results of sampling reference sites to set minimum criteria index scores for use designations in water quality standards. Provisions addressing biological criteria are in paragraph C of rule of the OAC. The antidegradation provisions describe the conditions under which water quality may be lowered in surface waters. Existing beneficial uses must be maintained and protected. Further, water quality better than that needed to protect existing beneficial uses must be maintained unless lower quality is deemed necessary to allow important economic or social development existing beneficial uses must still be protected. Provisions addressing

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antidegradation are in rule and rule of the OAC. Any interested individuals can have a role in the process of developing water quality standards. Ohio EPA reviews and, as appropriate, revises water quality standards at least once every three years. When water quality standards revisions are proposed, the public is notified of these revisions. A public hearing is held to gather input and comments. Summary Tables These tables summarize aquatic life and human health numerical water quality criteria and values contained in and developed pursuant to Chapter of the Ohio Administrative Code. The Lake Erie table was last updated on Feb. The Ohio River table was last updated on February 3,

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Chapter 2 : Ecological Risk Assessment Screening Benchmarks | Region 5 Superfund | US EPA

This project examined the development of ambient water quality criteria (AWQC) for the protection of wildlife for mercury. Mercury is considered a serious risk to wildlife in many areas. As a result, the Great Lakes Water Quality Initiative and others have developed AWQC.

Environmental Protection Agency, and approved for publication. Mention of trade names or commercial products. This document is a revision of proposed criteria based upon a consideration of comments received from other Federal agencies, State agencies, special interest groups, and individual scientists. The criteria contained in this document replace any previously published EPA aquatic life criteria. The term "water quality criteria" is used in two sections of the Clean Water Act, section 1 and section 2. The term has a different program impact in each section. In section 1, the term represents a non-regulatory, scientific assessment of ecological effects. The criteria presented in this publication are such scientific assessments. Such water quality criteria associated with specific stream uses when adopted as State water quality standards under section 2 become enforceable maximum acceptable levels of a pollutant in ambient waters. The water quality criteria adopted in the State water quality standards could have the same numerical limits as the criteria developed under section 1. However, in many situations States may want to adjust water quality criteria developed under section 2 to reflect local environmental conditions and human exposure patterns before incorporation into water quality standards. It is not until their adoption as part of the State water quality standards that the criteria become regulatory. Guidelines to assist the States in the modification of criteria presented in this document, in the development of water quality standards, and in other water-related programs of this Agency, have been developed by EPA. Acute Toxicity of Mercury to Aquatic Animals 26 2. Chronic Toxicity of Mercury to Aquatic Animals 36 3. Toxicity of Mercury to Aquatic Plants 44 5. Bioaccumulation of Mercury by Aquatic Organisms 46 6. Research Council Canada, ; Nriagu, Elemental mercury is a heavy liquid at room temperature and was considered relatively inert, because it was assumed that it would quickly settle to the bottom of a body of water and remain there in an innocuous state. However, elemental mercury is oxidized to mercury II under natural conditions Wood, Furthermore, mercury II, whether discharged directly or produced from elemental mercury, can be methylated by both aerobic and anaerobic bacteria Akagi, ; Beijer and Jernelov, ; Callahan, et al. Mercury II can also be methylated in the slime coat, liver, and intestines of fish Jernelov, ; Matsumura, et al. The term "methylmercury" is used herein to refer only to monomethylmercury, and not to dimethylmercury or any other monoorganomercury salt or diorganomercury compound. Inorganic mercury II will be referred to as "mercury II". The importance of methylation may be reduced by demethylation Bisogni, ; Ranaamorthy, et al. Demethylation also occurs in fish Burrows and Krenkel, ; de Freitas, et al. Numerous factors such as alkalinity, ascorbic acid, chloride, dissolved oxygen, hardness, organic complexing agents, pH, sediment, and temperature probably affect the acute and chronic toxicity and bioaccumulation of the various forms of mercury Amend, et al. A variety of studies have been conducted on the effect of selenium on the acute toxicity of mercury. Available data do not, however, show that quantitative relationships are consistent enough for a variety of aquatic species to enable relating water quality criteria to any of these variables. Because of the variety of forms of inorganic and organic mercury and lack of definitive information about their relative toxicities, no available analytical measurement is known to be ideal for expressing aquatic life criteria for mercury. Previous aquatic life criteria for mercury U. EPA, were specified in terms of total recoverable mercury, which would probably be measured as total mercury U. EPA, but both of these measurement methods are probably too rigorous in some situations. Acid-soluble mercury operationally defined as the mercury that passes through a 0.45 µm filter is compatible with all available data concerning toxicity of mercury to, and bioaccumulation of mercury by, aquatic organisms. No such results were rejected just because it was likely that they would have been substantially different if they had been reported in terms of acid-soluble mercury. For example,

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results reported in terms of dissolved mercury would not have been used if the concentration of precipitated mercury was substantial. On samples of ambient water, measurement of acid-soluble mercury should measure all forms of mercury that are toxic to aquatic life or can be readily converted to toxic forms under natural conditions. In addition, this measurement should not measure several forms, such as mercury that is occluded in minerals, clays, and sand or is strongly sorbed to particulate matter, that are not toxic and are not likely to become toxic under natural conditions. Although this measurement and many others will measure soluble, complexed forms of mercury, such as the EDTA complex of mercury II, that probably have low toxicities to aquatic life, concentrations of these forms probably are negligible in most ambient water. Although water quality criteria apply to ambient water, the measurement used to express criteria is likely to be used to measure mercury in aqueous effluents. Measurement of acid-soluble mercury should be applicable to effluents because it will measure precipitates, such as carbonate and hydroxide precipitates of mercury II, that might exist in an effluent and dissolve when the effluent is diluted with receiving water. If desired, dilution of effluent with receiving water before measurement of acid-soluble mercury might be used to determine whether the receiving water can decrease the concentration of acid-soluble mercury because of sorption. The acid-soluble measurement should be useful for resource needs, thus minimizing the number of samples and procedures that are necessary. The acid-soluble measurement does not require filtration at the time of collection, as does the dissolved measurement. Durations of 10 minutes to 24 hours between acidification and filtration probably will not affect the result substantially. Differences in pH within the range of 1. After acidification and filtration of the sample to isolate the acid-soluble mercury, the analysis for total acid-soluble mercury can be performed using permanganate and persulfate oxidation and cold vapor atomic absorption spectrophotometry (CVAAS), as with the total measurement. Methylmercury has been measured using gas chromatography (GC) (Cappon, ; Hildebrand, et al. Thus, expressing aquatic life criteria for mercury in terms of the acid-soluble measurement has both toxicological and practical advantages. On the other hand, because no measurement is known to be ideal for expressing aquatic life criteria for mercury or for measuring mercury in ambient water or aqueous effluents, measurement of both total acid-soluble mercury and total mercury in ambient water or effluent or both might be useful. For example, there might be cause for concern if total mercury is much above an applicable limit, even though total acid-soluble mercury is below the limit. Unless otherwise noted, all concentrations reported herein are expected to be essentially equivalent to acid-soluble mercury concentrations. All concentrations are expressed as mercury, not as the chemical species. The criteria presented herein supersede previous aquatic life water quality criteria for mercury (U.S. EPA,), because these new criteria were derived using improved procedures and additional information. Whenever adequately justified, a national criterion may be replaced by a site-specific criterion (U.S. EPA, b), which may include not only site-specific criterion concentrations (U.S. EPA, c), but also site-specific durations of averaging periods and site-specific frequencies of allowed exceedences (U.S. EPA, d). The literature search for information for this document was conducted in May, ; some newer information was also used. The latter information exists principally because many of these compounds were considered for use in treatment of diseases and control of parasites in fish culture, although their source for environmental concern is from industrial and agricultural uses for fungus control. Both phenylmercuric acetate and pyridylmercuric acetate have been called PMA. Tests have been conducted on different formulations which contain various percentages of active ingredient and the percentages of active ingredient given by the authors were used to calculate mercury concentrations. When the percentage of active ingredient was not given for pyridylmercuric acetate, 80 percent was assumed (Allison,). The freshwater acute toxicity values indicate that the difference in sensitivity between different species to a particular mercury compound is far greater than the difference in sensitivity of a particular species to various mercury compounds. For example, the reported acute values for mercury II range from 2. MacLeod and Pessah studied the effect of temperature on the acute toxicity of mercuric chloride to rainbow trout. Clemens and Sneed (b) found a similar effect of temperature on toxicity to 6 juvenile channel catfish; at 10°C, Acute values are available for more than one species in each of two genera, and the range of Species Mean

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Acute Values within each genus is less than a factor of 1. On the other hand, a midge was among the most sensitive species, whereas other insects were the most resistant species. The most sensitive genus, *Daphnia*, is times more sensitive than the most resistant, *Acroneuria* Table 3. A freshwater Final Acute Value of 4. Not enough data are available to calculate a Final Acute Value for methylmercury, but the available data indicate that it is more acutely toxic than mercury II. Saltwater fishes and invertebrates both show wide ranges of sensitivities to mercury II. Among invertebrates a mysid has an acute value of 3. Of the 29 saltwater genera for which acute values are available, the most sensitive, *Mysidopsis*, is times more sensitive than the most resistant, *Pseudopleuronectes*. Acute values are available for more than one species in each of three genera and the range of Species Mean Acute Values within each genus is less than a factor of 1. The saltwater Final Acute Value of 4. In addition, a chronic test with brook trout on methylmercuric chloride yielded a chronic value of 0. Both an early life-cycle test and a life-cycle test on mercuric chloride found adverse effects on the fathead minnow at all concentrations tested including the lowest of 0. For mercuric chloride the acute-chronic ratio with *Daphnia magna* is less than 6, whereas that with the fathead minnow is greater than 6. For methylmercury the acute-chronic ratio with brook trout is 1. No spawning occurred at 2. Time to spawn and productivity at 1. The highest concentration at which no statistically significant effect on reproductive processes was detected was 0. Therefore, the chronic limits are 0. The 96 hr LC50 for this species in the same study was 3. The species mean acute-chronic ratio for *Daphnia magna* is 4. These are sensitive species in fresh and salt water, respectively, and the four most sensitive species in each 8 Bioaccumulation Bioconcentration is a function of the relative rates of uptake and depuration. The bioconcentration factor BCF of mercury is high for fish because uptake is relatively fast and depuration is very slow. Thus, the biological half-life of mercury in fish is approximately 2 to 3 years de Freitas, et al. Depuration of mercury is so slow that, even in the absence of exposure to mercury, long-term reduction in the concentration of mercury in fish tissue is largely due to dilution by tissue addition from growth. Usually less than 60 percent of mercury in invertebrates is methylated, but in fish, except for young fish, usually more than 70 percent is methylated Bache, et al. The distribution of mercury within a fish is the result of the movement of mercury from the absorbing surfaces gills, skin, and gastrointestinal tract, into the blood, then to the internal organs, and eventually either to the kidney or bile for recycling or elimination or to muscle for long-term storage. As the tissue concentration approaches steady-state, net accumulation rate is slowed either by a reduction in uptake rate, possibly due to inhibition of membrane transport, or by an increase in depuration rate, possibly because of a saturation of storage sites, or both. High concentrations of mercury in the slime coat of certain freshwater fishes, such as burbot, eels, and northern pike, and in the skin of acutely-exposed fishes are believed to be due to the methylating activity of bacteria prevalent in the mucous coat Jernelov, In addition, acutely toxic concentrations of mercury have been reported to stimulate secretion of mucus Baker, ; Lock, et al.

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Chapter 3 : Ambient Water Quality Criteria for Mercury

This table contains the most up to date criteria for aquatic life ambient water quality criteria. Aquatic life criteria for toxic chemicals are the highest concentration of specific pollutants or parameters in water that are not expected to pose a significant risk to the majority of species in a.

Environmental Protection Agency, and approved for publication. Mention of trade names or commercial products does not constitute endorsement or recommendation for use. Proposed water quality criteria for the 65 toxic pollutants listed under section a 1 of the Clean Water Act were developed and a notice of their availability was published for public comment on March 15, 44 FR , July 25, 44 FR , and October 1, 44 FR. This document is a revision of those proposed criteria based upon a consideration of comments received from other Federal Agencies, State agencies, special interest groups, and individual scientists. The criteria contained in this document replace any previously published EPA criteria for the 65 pollutants. This criterion document is also published in satisfaction of paragraph 11 of the Settlement Agreement in Natural Resources Defense Council, et. The term "water quality criteria" is used in two sections of the Clean Water Act, section a 1 and section c 2. The term has a different program impact in each section. In section , the term represents a non-regulatory, scientific assessment of ecological effects. The criteria presented in this publication are such scientific assessments. Such water quality criteria associated with specific stream uses when adopted as State water quality standards under section become enforceable maximum acceptable levels of a pollutant in ambient waters. The water quality criteria adopted in the State water quality standards could have the same numerical limits as the criteria developed under section However, in many situations States may want to adjust water quality criteria developed under section to reflect local environmental conditions and human exposure patterns before incorporation into water quality standards. It is not until their adoption as part of the State water quality standards that the criteria become regulatory. Guidelines to assist the States in the modification of criteria presented in this document, in the development of water quality standards, and in other water-related programs of this Agency, are being developed by EPA. Environmental Protection Agency John H. Environmental Protection Agency Leonard T. Environmental Protection Agency Rudy J. For total recoverable mercury the criterion to protect saltwater aquatic life as derived using the Guidelines is 0. Criteria reflect ingestion of marine organisms as well as fresh-water and estuarine fish and shellfish. Mercury is a silver-white metal, atomic weight A liquid at room temperature, its melting point is The metal is insoluble and is not attacked by water. Mercury exists in a number of forms in the environment. Mercurous salts are much less soluble in water. HgN03 will solubilize only in 13 parts water containing 1 per- cent HN Hg2Cl2 is practically insoluble in water. Because of this, mercurous salts are much less toxic than the mercuric forms Stecher, The Department of the Interior carried out a nationwide reconnaissance of mercury in U. The higher mercury concentrations were generally found in small streams. About half of the 43 samples from the Mississippi River contained less than 0. The mercury content of lakes and reservoirs was between 0. With few exceptions, the mercury content of groundwater samples was below detection 0. Of these, or Eleven of the supplies had mercury concentrations of 1. When this one supply was extensively reexamined, the mercury concentration was found to be less than 0. In a study of Pacific waters, mercury concentrations were found to increase from surface values of about 0. In an area seriously affected by pollution Minamata Bay, Japan , values ranged from 1. The National Research Council has shown typical oceanic values for mercury to be 0. A major use of mercury has been as a cathode in the electrolytic preparation of chlorine and caustic soda; this accounted for 33 percent of total demand in the United States in Electrical apparatus lamps, arc rectifiers, and mercury battery cells accounted for 27 percent, industrial and control instruments switches, thermometers, and barometers , and general laboratory applications accounted for 14 percent of demand. Use of mercury in antifouling and mildew proofing paints 12 percent and mercury formulations used to control fungal diseases of seeds, bulbs, plants, and vegetation 5 percent were other major utilizations, however,

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mercury is no longer registered by the EPA for use in antifouling paints or for the control of A-2 fungal diseases of bulbs. The remainder 9 percent was for dental amalgams, catalysts, pulp and paper manufacture, Pharmaceuticals, and metallurgy and mining. Several forms of mercury, ranging from elemental to dissolved inorganic and organic species, are expected to occur in the environment. The finding that certain microorganisms have the ability to convert inorganic and organic forms of mercury to the highly toxic methyl or dimethyl mercury has made any form of mercury highly hazardous to the environment Jensen and Jernelov, In water, under naturally occurring conditions of pH and temperature, inorganic mercury can be converted readily to methyl mercury Bisogni and Lawrence, Mercury is able to form a series of organometallic compounds with alkyl, phenyl, and methoxyethyl radicals. Methylation of mercury in aerobic and anaerobic environments. Center, Ithaca, New York. Mercury in drinking water supplies. Mercury in waters of the United States, Biological methylation of Nature. An assessment of mercury in the environment. Mercury in the environment: Elemental mercury, which is a heavy liquid at room temperature, was considered relatively inert. It was thought that it would quickly settle to the bottom of a body of water and remain there in an innocuous state. However, elemental mercury can be oxidized in sediment to divalent mercury Wood, Furthermore, both aerobic and anaerobic bacteria have been found capable of methylating divalent mercury in sediments The National Research Council, and estuarine areas Jernelov, This methylated form is more water soluble and more biologically active than elemental and inorganic divalent mercury Fromrn, ; Armstrong and Scott ; Jernelov, et al. Largely because of bacterial methylation, mercury is much more of a serious threat to the aquatic environment than was suspected. Mercury is one of the few pollutants that, at about the same concentrations in water, adversely affects aquatic life through direct toxicity and affects uses of aquatic life through bioaccumulation. Bioaccumulation has received more attention because of potential adverse effects to humans. Methylmercury is more toxic than inorganic mercury to mammals as well as aquatic life, and mercury has no known physiological function. The following tables contain the appropriate data that were found in the literature, and at the bottom of each table are calculations for deriving various measures of toxicity as described in the Guidelines. Even in situations in which no organic mercury was known to have been discharged, methylmercury was the dominant form in tissue residues Jernelov and Lann, The methylated form is of great concern because it comprises most of the mercury residue in tissues of aquatic organisms Hattula, et al. Defining the toxicity of mercury residues to humans, and probably other consumers of aquatic life, is complicated by the effect of selenium on the toxicity of mercury Strom, et al. The FDA action level of 1. Once methylation takes place, uptake by aquatic life is extremely rapid, and demethylation is a very slow process McKim et al. Deuration by excretion through the kidney reportedly requires demethylation Burrows and Krenkel, In freshwater fishes, initial elimination of mercury just after the end of exposure is relatively rapid, due to sluffing of the slime coat Burrows and Krenkel, and elimination of non-methylated mercury. Once methylmercury becomes securely bound to sulfhydryl groups in muscle proteins, subsequent loss proceeds at a much reduced rate. In fact, long term reduction of the concentration of mercury in fish tissue is largely due to dilution by tissue addition resulting from growth Lockhart, et al. High mercury concentrations in slimy freshwater fishes such as burbot, eels, and northern pike, and in the skin of acutely-exposed fishes are believed due to the methylating activity of bacteria prevalent in the mucous coat Jernelov, Acutely toxic concentrations of mercury have been reported to stimulate mucous secretion McKone, et al. However, these are the layers of the fish that are first encountered as mercury moves from the environment into a fish, and in acute exposures the mercury does not have time to be transported to the final sink - the proteins whose greatest mass are in the axial muscle McKim, et al. Numerous data are available concerning the effect of phenylmercuric acetate PMA on aquatic organisms, because of its use as a fungicide and its use to treat fish diseases. Many tests have been conducted on different PMA formulations which contain various percentages of active ingredient. The percentages of active ingredient given by the authors were used to convert to concentrations of mercury. When the percentage of active ingredient was not given, 80 percent PMA was assumed Allison, Of the analytical measurements currently available, water quality criteria for mercury are

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probably best stated in terms of total recoverable mercury, because of the variety of forms of mercury that can exist in bodies of water and the various chemical and toxicological properties of these forms. The forms of mercury that are commonly found in bodies of water and are not measured by the total recoverable procedure, such as the mercury that is a part of minerals, clays and sand, probably are forms that are less toxic to aquatic life and probably will not be readily converted to the more toxic forms under natural conditions. On the other hand, forms of mercury that are commonly found in bodies of water and are measured by the total recoverable procedure, such as the free ion, the hydroxide, carbonate, and sulfate salts, and the organic compounds, probably are forms that are more toxic to aquatic life or can be converted to the more toxic forms under natural conditions. Because the criteria for mercury are derived on the basis of tests conducted on soluble inorganic salts of divalent inorganic mercury and monomethylmercuric chloride, the total and total recoverable concentrations in the test should be about the same. Except as noted, all concentrations reported herein are expected to be essentially equivalent to total recoverable mercury. All concentrations are expressed as mercury, not as the compound. The latter information exists principally because many of these compounds have been used for disease treatment and parasite control in fish cultural practices, though their source for environmental concern is from industrial and agricultural uses for fungus control. A striking feature of the freshwater acute toxicity values is that the difference in sensitivity between different types of organisms to a particular mercury compound is far greater than the difference in sensitivity of a particular species to various mercury compounds. For inorganic mercury, the reported hour LC50 values range from 0. Data are insufficient to make such comparisons for other two classes of mercury compounds. Rainbow trout are the most acutely sensitive of the tested fish species to all three kinds of mercury compounds and methylmercuric chloride is about ten times more acutely toxic to rainbow trout than is mercuric chloride. MacLeod and Pessah studied the effect of temperature on the acute toxicity of mercuric chloride to rainbow trout. Clemens and Sneed found similar temperature effects with mercury exposures at 10, A freshwater Final Acute Value of 0. This value should be useful because it is based on data for eleven species, even though acute data are not available for any non-salmonid fish. Acute values for mercuric chloride are available for 26 species of saltwater animals from 5 phyla Table 1. Fishes were generally more resistant to mercuric chloride than the crustaceans and molluscs. The saltwater Final Acute Value for inorganic mercury, derived from the species mean acute values in Table 3 using the calculation procedures described in the Guidelines is 3. Chronic Toxicity Chronic toxicity tests with *Daphnia magna* have been conducted on three different kinds of mercury compounds and the chronic values were all between 1.

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Chapter 4 : Water Quality Standards

Overview This project examined the development of ambient water quality criteria (AWQC) for the protection of wildlife for mercury. Mercury is considered a serious risk to wildlife in many areas.

PAHs, additional pesticides, additional inorganics The Ecological Soil Screening Level Eco-SSL derivation process represents the collaborative effort of a multi-stakeholder workgroup consisting of federal, state, consulting, industry and academic participants led by the U. It is emphasized that the Eco-SSLs are soil screening numbers, and as such are not appropriate for use as cleanup levels. Screening ecotoxicity values are derived to avoid underestimating risk. Requiring a cleanup based solely on Eco-SSL values would not be technically defensible. The Eco-SSL web site provides an overview of the contaminant. Separate discussions are provided for each receptor group including a comprehensive list of literature evaluated under the effort, and a summary of data used in deriving Eco-SSL values. Calculation and evaluation of sediment effect concentrations for the amphipod *Hyaella azteca* and the midge *Chironomus riparius*. Suter II, and R. The representative effect concentration selected from among the high no-effect-concentrations for *Hyaella azteca* and *Chironomus riparius* are presented in EPA based on the ranking method presented in Jones et al. The majority of the data are for freshwater sediments. These values facilitate consistency in screening level ecological risk assessments throughout Region III. Marine and freshwater sediment, surface water Contaminants: PCBs and pesticides; PAHs and other organics; inorganics The tables include compounds for which benchmark values have been established or that are considered bioaccumulative compounds identified in tables. Region 4 Bulletins No. Office of Health Assessment. Values presented are as updated Aug. Soil, sediment, surface water Contaminants: PCBs and pesticides; PAHs and other organics; inorganic The Region 4 surface water screening values were obtained from Water Quality Criteria documents and represent the chronic ambient water quality criteria values for the protection of aquatic life. These are possible effects benchmarks. They are not intended to serve as cleanup levels. Army Corps of Engineers U.

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This project examined the development of ambient water quality criteria (AWQC) for the protection of wildlife for mercury. Mercury is considered a serious risk to wildlife in many areas.

Contact Us Fact Sheet: The Ambient Water Quality Criteria National Guidelines outline the methodology used by states and tribes to develop human health water quality criteria. Revisions to the guidelines incorporate significant scientific advances in key areas such as cancer and non-cancer risk assessments, exposure assessments, and bioaccumulation in fish. The revised methodology will provide more flexibility for decision-making at the state, tribal and EPA regional levels. It is most likely that the methodology will result in more stringent criteria for bioaccumulatives and generally similar values of nonbioaccumulatives. Human health criteria are developed under Section a of the Clean Water Act of and are designed to protect human health. Water quality criteria are developed by assessing the relationship between pollutants and their effect on human health and the environment. These criteria are used by states and Indian tribes to establish water quality standards and ultimately provide a basis for controlling discharges or releases of pollutants. These national guidelines addressed three types of endpoints: The states and tribes use these criteria to develop water quality standards for each water body. EPA is required to review periodically criteria adopted by states and tribes. They provide detailed means for developing water quality criteria, including systematic procedures for evaluating cancer risk, noncancer health effects, human exposure, and bioaccumulation potential in fish. EPA Methodology for Deriving Criteria States and tribes must develop water quality standards that include designated uses and water quality criteria necessary to support those uses. The Methodology is the guidance for states and tribes to help them establish water quality criteria and standards to protect human health. It provides detailed means for developing the water quality criteria, including systematic procedures for evaluating cancer risk, noncancer health effects, human exposure, and bioaccumulation potential in fish. Risk assessment practices have evolved significantly since , particularly in the areas of cancer and noncancer risk assessments with new information, procedures, and numerous published Agency guidelines , exposure assessments with new studies on human intake and exposure patterns, and new science policy guidelines and methodologies on accounting for bioaccumulation in fish. General Background of the Revision Process Revisions began with a national workshop in , where participants discussed critical issues. Based on individual expertise, attendees were assigned to technical workgroups including cancer risk, noncancer risk, exposure, and bioaccumulation in fish. EPA created a workgroup in , including program office and regional participants, to revise the methodology. Following publication of the draft Methodology revisions, written public comments were accepted. In May , a peer review workshop was held. A public stakeholders meeting was also held then. EPA received extensive input on the Methodology from each of these groups. Major Methodology Revisions Publication of final revisions satisfies the requirements of the CWA that EPA periodically revise criteria for water quality to reflect accurately the latest scientific knowledge on the kind and extent of all identifiable effects on health and welfare that may be expected from the presence of pollutants in any body of water. These Final AWQC Methodology Revisions to the AWQC National Guidelines are necessitated by the many significant scientific advances made during the past 20 years in the key areas of cancer and noncancer assessments, exposure assessments, and bioaccumulation in fish. The major revisions are in four assessment areas: For carcinogen cancer risk assessment: Recommend more sophisticated methods to comprehensively determine the likely mechanism of human carcinogenicity. Recommend a mode of action MOA approach to determine the most appropriate low- dose extrapolation for carcinogenic agents. Recommend consideration of other issues related to the RfD process including: Recommend the use of quantitative dose-response modelling for the derivation of RfDs. Provide guidance for states and tribes on the use of an alternative value from the RfD point estimate, within a limited range, to reflect the inherent imprecision of the RfD. Encourage states and tribes to use local studies on fish consumption that better reflect local intake patterns and choices.

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Recommend default fish consumption values for the general population, recreational fishers and subsistence fishers. Account for other sources of exposure, such as food and air, when deriving AWQC for noncarcinogens and nonlinear carcinogens. Focus on the use of bioaccumulation factors BAFs, instead of bioconcentration factors BCFs for estimating potential human exposure to contaminants via the consumption of contaminated fish and shellfish. Use high quality field data over laboratory or model-derived estimates for deriving BAFs, since field data best reflect factors which can affect the extent of bioaccumulation. EPA does not plan to completely revise all of the criteria developed in or those updated as part of the National Toxics Rule. Partial updates of all criteria may be necessary. EPA will continue to develop and update toxicology and exposure data needed in the derivation of AWQC that may be impractical for the states and regions to obtain. The development of revised criteria for chemicals of high priority and national importance including, but not limited to, chemicals that bioaccumulate, such as PCBs, dioxin, and mercury. The development or revision of AWQC for some additional priority chemicals. EPA encourages states and tribes to use the revised methodology to develop or revise AWQC to reflect local conditions appropriately. EPA believes that AWQC inherently require several risk management decisions that are, in many cases, better made at the state and regional level. EPA believes the AWQC require several risk management decisions that are often better made at the state, tribal and regional level. The methodology will probably result in more stringent criteria for bioaccumulatives due to the use of BAFs instead of BCFs and generally similar, or less stringent, values of nonbioaccumulatives. How do I get a copy of this technical support document?

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Chapter 7 : State of Oregon: Water Quality - Water Quality Standards

ambient aquatic life water quality criteria for mercury u.s. environmental protection agency office of research and development environmental research laboratories duluth, minnesota narragansett, rhode island.

Chapter 8 : Developing Ambient Water Quality Criteria for Mercury

EPA's Methodology for Deriving Ambient Water Quality Criteria for the Protection of Human Health is used in the development of EPA's recommended criteria and offered as guidance for states and tribes in developing their own criteria.

Chapter 9 : ambient water quality criteria recommendations | Download eBook pdf, epub, tuebl, mobi

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